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Habitat ecology of Nearctic–Neotropical migratory landbirds on the nonbreeding grounds

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ABSTRACT

Of the approximately 716 bird species that breed in North America, 386 (54%) are considered Nearctic–Neotropical migrants by the U.S. Fish and Wildlife Service. In the past 50 yr, scores of these migratory species, including some once considered common, have declined dramatically. Nearctic–Neotropical migrants normally spend 6–8 months in tropical habitats, making the identification, availability, and management of Neotropical habitats critical issues for their conservation. Yet, for most species, complete and nuanced information about their use of tropical habitats and the relative effects of breeding vs. wintering conditions on survival, productivity, and population trends is not available, though many studies point to Neotropical overwintering habitats as being a strong driver of population change. Particularly important for long-distance Nearctic–Neotropical migrants is an understanding of how “carry-over effects” arise and influence population trends when conditions on wintering grounds and tropical stopover areas affect subsequent reproductive performance on breeding grounds. For example, why some species show strong carry-over effects from tropical habitats while others do not is not fully understood. In recent years, many studies have offered insights into these issues by taking advantage of new scientific methods and technological innovations. In this review, we explore threats facing North American breeding birds that migrate to the Neotropics, summarize knowledge of habitat selection and use on the wintering grounds, describe how conditions at one point in the annual cycle may manifest in subsequent seasons or life history stages, and discuss conservation concerns such as climate change and the potential for phenological mismatch.

Keywords: carry-over effects, landbirds, migration, Nearctic–Neotropical migrants, phenological mismatch, wintering habitat

LAY SUMMARY

- More than half of the bird species that breed in North America, representing billions of birds, migrate to the Neotropics.
- In the past 50 years, scores of these species have declined dramatically.
- Migratory birds normally spend 6–8 months in tropical habitats, making the conservation of these species an international challenge. Yet, for most species, complete and nuanced information about their use of tropical habitats and the relative effects of breeding vs. wintering conditions on survival, productivity, and population trends is not available.
- Accelerating climate change is adding to the urgency of our gaining an understanding of the full annual and migratory cycle of these birds.
- In recent years, many studies have offered insights into these issues by taking advantage of new scientific methods and technological innovations.

Ecología del hábitat de las aves terrestres migratorias Neártico–Neotropicales en las zonas no reproductivas

RESUMEN

De las aproximadamente 716 especies de aves que se reproducen en América del Norte, el Fish and Wildlife Service de los EE. UU. considera que 386 (54%) son migratorias neártico-neotropicales. En los últimos 50 años, decenas de estas

especies migratorias, incluidas algunas que hace poco se consideraban comunes, han disminuido drásticamente. Las aves migratorias neártico-neotropicales normalmente pasan entre seis y ocho meses en hábitats tropicales, lo que hace que la identificación, disponibilidad y manejo de los hábitats neotropicales sean temas críticos para su conservación. Sin embargo, para la mayoría de las especies, no se dispone de información completa y detallada sobre el uso de hábitats tropicales y los efectos relativos de las condiciones de reproducción vs. invernada en la sobrevivencia, la productividad y las tendencias de la población, aunque muchos estudios señalan que los hábitats de invernada neotropicales son un factor determinante de los cambios poblacionales. Es particularmente importante para los migrantes neárticos-neotropicales de larga distancia, es la comprensión de cómo surgen los “carryover effects” e influyen en las tendencias de la población cuando las condiciones en las zonas de invernada y en las zonas de escala tropical afectan el rendimiento reproductivo posterior en las zonas de reproducción. No se comprende completamente, por ejemplo, por qué algunas especies muestran fuertes “carry-over effects” desde los hábitats tropicales, mientras que otras no. En los últimos años, muchos estudios han ofrecido nuevas perspectivas sobre estas cuestiones aprovechando los nuevos métodos científicos e innovaciones tecnológicas. En este capítulo exploramos las amenazas a las que se enfrentan las aves reproductoras de América del Norte que migran al Neotrópico, resumimos el conocimiento del uso y la selección del hábitat en sus zonas de invernada, describimos cómo las condiciones en un punto del ciclo anual pueden manifestarse en las siguientes estaciones o etapas de la historia de la vida y discutimos problemas de conservación como el cambio climático y el potencial de desajuste fenológico.

Palabras clave: aves terrestres, carry-over effects, hábitat de invernada, migración, , migratorios Neártico–Neotropicales

INTRODUCTION

From a physiological perspective, long-distance migration is taxing (McWilliams and Karasov 2014). Birds must spend days or weeks gaining and storing energy to fly hundreds or thousands of kilometers over oceans, mountains, or deserts while avoiding adverse weather, predators, food shortages, and anthropogenic obstacles such as those posed by development (Hedenström 2008, Newton 2008, Longcore et al. 2013; Figure 1). When migrants arrive in the Neotropics, they may have to compete with conspecifics for resources, relegating less competitive individuals to lower quality habitat (Marra et al. 1993, Marra 2000, Studds and Marra 2005). Despite these challenges, the benefits of migration from the tropics to temperate zones (to exploit an abundance of spring and summer food resources) and back to the tropics (to escape harsh northern winter weather and lack of food) appear substantial, and these tradeoffs have shaped the remarkable life history strategies exhibited by Nearctic–Neotropical migratory birds (Alerstam 1990). While important knowledge about full annual cycle dynamics has increased greatly, our understanding of the specific drivers of population decline, especially in tropical wintering habitats, remains relatively poor (Runge and Marra 2005, Sherry et al. 2015).

This is no small matter: bird populations across the Americas are in trouble. A recent estimate from North America has put the loss at 25% of the continental avifauna—nearly 3 billion birds—in less than half a century (Rosenberg et al. 2019). Of ~716 bird species that breed in North America, 386 (54%) are considered Neotropical migrants by the U.S. Fish and Wildlife Service (U.S. Fish and Wildlife Service 2000), and migratory species might be particularly at risk compared with nonmigrants: Runge et al. (2015) noted that only 9% of migratory

species are adequately covered by protected areas across all stages of their annual cycle, in comparison with 45% of nonmigratory birds.

Until conservation professionals better understand full annual cycle dynamics, such as how factors related to breeding in the Nearctic compared with wintering in the Neotropics drive population change, effective conservation action will be difficult, as efforts to conserve species by increasing or enhancing habitats during the less important parts of the cycle will have little effect on population size (Sherry and Holmes 1995, Sheehy et al. 2010, Rushing et al. 2016). Specifically, if a species is declining due to factors operating on the wintering grounds in the Neotropics rather than the temperate breeding grounds, conservation efforts on wintering grounds are likely to be most effective, and vice versa. (Note: although the term “winter” may refer to different calendar periods in temperate and tropical latitudes, the terms “winter” or “wintering” as we use them in this paper are synonymous with the nonmigratory, nonbreeding season of Nearctic–Neotropical migratory birds.)

This is an exciting time to study Nearctic–Neotropical migratory birds. A growing body of Latin American professionals and an increase in international, collaborative studies have led to more and deeper field studies in the tropics. For example, multiple studies in northern Colombia have drawn attention to the role tropical habitats play in enabling birds to prepare for (in spring) and recover from (in fall) long migratory flights across the Caribbean Sea (Bayly et al. 2013, 2019, Gómez et al. 2015). Multi-country initiatives are also becoming more common, shedding light on patterns of connectivity and variation in abundance at scales rarely studied in the past (González-Prieto et al. 2016, Bulluck et al. 2019), and Neotropical agroecosystems such as shade



FIGURE 1. The full annual cycle of Nearctic–Neotropical migratory birds. Although a generalized representation, most birds spend the majority of the year in the Neotropics. (Artwork by Lauren Helton, The Institute for Bird Populations.)

coffee and silvo-pastures have received attention as researchers have learned more about their use as winter habitat for migrants (McDermott and Rodewald 2014, Colorado and Rodewald 2017, González-Prieto 2018b, Narango et al. 2019).

Simultaneously, several studies have highlighted the importance of migration and stopover habitats (Bayly et al. 2012, Wolfe et al. 2014), a life history phase that is recognized as contributing to a large portion of annual mortality (Sillett and Holmes 2002, Rockwell et al. 2017). A relative explosion in the use of eBird in the Neotropics has opened the door to a plethora of questions regarding nonbreeding distributions (La Sorte et al. 2017) and movement patterns throughout the region (La Sorte et al. 2016b); and recent advances in tracking technology,

such as miniaturized global positioning systems, light-level geolocators, and the Motus system (Taylor et al. 2016), are being used routinely to shed light on full annual cycle ecology and the connections between temperate and tropical habitats (Wikelski et al. 2007, Bridge et al. 2011, Taylor et al. 2016). Innovations in molecular genetics have also improved understanding of taxonomy and species distributions (Ruegg et al. 2014). Genome sequencing and isotopic signature study, among other techniques, have enabled subspecies or population differentiation as well as the linkage of breeding and wintering sites with migratory pathways and stopover sites (González-Prieto and Hobson 2013, Ruegg et al. 2014, Hobson et al. 2015b, González-Prieto et al. 2016, Ruegg et al. 2016).

Awareness of emerging and potentially grave threats has paralleled this rapid expansion of knowledge. Species declines due to habitat loss and fragmentation by anthropogenic causes are accelerating in many parts of the Neotropics (Rodríguez-Eraso et al. 2013, Ceballos et al. 2015), and a rapidly changing climate with potentially shifting bottom-up trophic effects that may affect the availability and phenology of food resources adds several layers of complexity. Thus, while advances in ornithology and technology have produced useful insights into the full annual cycle of Nearctic–Neotropical migratory birds, important questions remain, and the context for which bird conservation will be carried out in the 21st century and beyond is rapidly shifting. Recent findings by Rosenberg et al. (2019) have created renewed attention in the plight and loss of migratory bird populations, making a reexamination of the state of our knowledge timely. In this review, we aim to highlight important knowledge gaps and identify future directions for Neotropical migratory bird research and conservation.

Habitat Conservation Concerns and Threats

With the worldwide human population already exceeding 7.6 billion and projected to reach 11 billion by the end of the 21st century, threats to birds will increase. Habitat is being lost, fragmented, and degraded due to the direct and indirect effects of an expanding agricultural footprint, climate change, environmental contaminants, invasive species, urban development, and myriad other threats. Forests in areas such as the Sierra Madre of Mexico and the heavily populated tropical Andes Mountains continue to be converted to other uses less sustainable to birds (Rodríguez-Eraso et al. 2013), threatening vulnerable species such as Cerulean Warbler (*Setophaga cerulea*), Wood Thrush (*Hylocichla mustelina*), Olive-sided Flycatcher (*Contopus cooperi*), and Canada Warbler (*Cardellina canadensis*), not to mention scores of (currently) more common species. Forest loss continues throughout much of the remaining large tracts of Central American wet and dry tropical forests, putting many migratory species at risk (Blandón et al. 2016). Meanwhile, the acceleration of forest habitat loss in the Amazon may affect populations of so far little known core wintering habitat of forest-dependent species such as Veery (*Catharus fuscescens*; Remsen 2001) and Connecticut Warbler (*Oporornis agilis*; Diniz et al. 2014). Other tropical habitats, including mangroves, which are important for many migratory warblers (Wunderle et al. 2014, Bulluck et al. 2019), and native grasslands (Vickery et al. 1999), are now rare in both North and South America, and are rapidly disappearing. However, certain types of land conversion of forest to agriculture or other uses may not be universally negative for Nearctic–Neotropical migratory birds. Some studies have shown increased avian

diversity in mixed agroforestry landscapes (Bakermans et al. 2009, DeBeenhouwer et al. 2013, González-Prieto 2018b, Valdéz-Juárez 2018), although these benefits appear to accrue largely for habitat generalists. In addition, research has emphasized the potential for population-level impacts to occur during migration, especially at stopover sites (Baker et al. 2004, Faaborg et al. 2010, Cohen et al. 2015). Stopover habitat is being lost at similar rates to wintering habitat but, due to concentration effects at migratory bottlenecks, the impacts may be more widespread and affecting multiple populations. In general, these areas have received much less scientific and conservation attention than breeding and wintering areas.

Understanding Habitat Quality

Determining which conservation strategies work best for Nearctic–Neotropical migratory birds requires an understanding of what constitutes habitat quality. Conservation of migratory birds in the Neotropics has tended to focus on mitigating habitat loss (i.e. quantity), but habitat quality also merits discussion, and progress is ongoing (e.g., Johnson 2007, Faaborg et al. 2010). From studies in Jamaica on American Redstarts (*Setophaga ruticilla*), Johnson et al. (2006a) found that birds survive best from one winter to the next in habitats where they maintain body mass over the course of the dry season. Peele (2015) found that moderate elevation wet limestone forest had the highest estimates of annual survival, relatively early departure for spring migration, proportionately older males, and a near absence of transient individuals, all of which are indicators of high habitat quality. Early departure time in migration contributes to early arrival on the breeding grounds and enhanced reproductive success (Reudnik et al. 2009, Cooper et al. 2015). Shade coffee plantations are another higher elevation woodland habitat that provides abundant food and can constitute a refuge for wintering migratory birds in the Caribbean and Central America (see Wunderle and Arendt 2017). In a different context, Ruiz-Sánchez et al. (2017) found higher densities and smaller territory sizes for wintering Wilson's Warblers (*Cardellina pusilla*) in conserved cloud forests in southeastern Mexico compared with unprotected sites with lower forest cover, but without finding differences in body condition. This difference between habitat types suggests that density was mediated by food availability, with birds reducing territory size as resources increased. A similar finding was found in work on Swainson's Thrush (*Catharus ustulatus*) wintering in forest vs. shade coffee in Colombia (González-Prieto 2018b).

Body condition, which is often measured using body mass maintenance or gain, has been widely used to assess winter habitat quality (Smith et al. 2010, Colorado and Rodewald 2017), including during migration stopovers

(Buler et al. 2007, Bayly et al. 2016). More research is needed to understand precisely how to estimate winter body condition and its components in physiologically and ecologically meaningful ways.

In a comparison of stopover habitats in Colombia, Bayly et al. (2016) examined 7 potential parameters as possible indicators of habitat quality. Rate of body-mass change, foraging rate, recapture rate, bird density, and flock size were expected to show a positive correlation with habitat quality. In addition, more adults and more males and birds of higher body mass might be expected in higher quality habitats because of their ability to competitively exclude younger, smaller, or less experienced conspecifics and deny them access to resources. The authors found higher densities, body mass, foraging rates, and rates of mass change in pre-montane forest than in shade-coffee plantation for Tennessee Warbler (*Leiothlypis peregrina*) but no clear evidence for differences in age or sex ratios. For Gray-cheeked Thrush (*Catharus minimus*), forest provided higher body mass and mass gains, while shade coffee provided poorer fueling conditions (measured by fewer recaptures). These results point to the need to use multiple, independent indicators that are relevant to each species' ecology when assessing habitat quality.

Carry-Over Effects

Poor quality tropical habitat can have long-term negative effects on subsequent phases of an individual bird's annual cycle, a process known as seasonal carry-over (e.g., Norris and Taylor 2006, Hostetler et al. 2015). In some dry tropical habitats, arthropod abundance can be lower in the dry season, which has been linked to lower body mass (Cooper et al. 2015). In American Redstarts, the most extensively studied species with respect to carry-over effects, individuals occupying territories in dry scrub (as opposed to mangrove) were in worse physiological condition prior to the onset of spring migration, leading to delayed arrival times on the breeding grounds, which is known to affect a migratory bird's fitness (Marra and Holberton 1998). A decline in body condition of ~10% over the dry season was associated with a delay of almost 1 week in spring migration departure (Studds and Marra 2007). Experimental reductions in food availability led to lowered body condition and delayed departure (Cooper et al. 2015), and carbon isotope analysis revealed that former winter habitat of individual redstarts arriving in spring linked higher winter habitat quality to earlier male arrival and higher reproductive success (Norris et al. 2004). Similar findings have been made for Kirtland's Warbler (*Setophaga kirtlandii*; Rockwell et al. 2012) and Barn Swallow (*Hirundo rustica*; Saino et al. 2004).

Other research has shown that birds arriving later at spring stopover or breeding sites had stable isotope signatures

associated with drier wintering habitat (Gonzalez-Prieto and Hobson 2013, Paxton and Moore 2015, Akresh et al. 2019). For Gray-cheeked Thrush on spring stopover at sites in northern Colombia, acquired fuel loads carried over influenced the subsequent pace of migration (less fuel meant a slower pace), and likely delayed arrival at the breeding grounds, possibly by weeks (Gómez et al. 2017).

Direct tracking studies confirm that, for several species, late departure from the wintering grounds correlates with late arrival at the breeding site (Stanley et al. 2012), and lab studies have shown a link between high energetic demands of flight endurance in females and subsequent lower quality of eggs (Skrip et al. 2016), either of which could reduce reproductive success.

By contrast, in Wood Thrush, a geolocator migration tracking study found that habitat quality at a wintering site in Belize and premigration body condition did not affect spring migration timing, speed, or arrival time at the breeding site (McKinnon et al. 2013). Wood Thrushes are relatively large-bodied and omnivorous, and individuals are consistent from year to year in their migration timing relative to the population average, suggesting an endogenous migration schedule (Stanley et al. 2012). Individuals in relatively poor condition may leave "on time" and then compensate at stopovers through differential habitat selection, diet, and/or duration of stopover. Some studies on smaller songbirds have also found that carry-over effects from tropical habitats can be weak. Research using stable isotopes for an Arctic-breeding population of Yellow Warbler (*Setophaga petechia*) did not detect any carry-over effects of nonbreeding habitat on arrival date or reproductive success (Drake et al. 2014). No carry-over effects were detected in one study of Magnolia Warblers (*Setophaga magnolia*) captured during spring migration (Boone et al. 2010). Thus, the American Redstart "model" may not apply equally well to all species, perhaps because of differences in body size, diet, habitat use, or migration strategy. Also, the gradients in tropical habitat quality occupied by some species may not be extreme enough to cause strong carry-over effects. For instance, body condition in Wilson's Warblers did not differ between a large protected cloud forest habitat and a small unprotected fragment (Ruiz-Sánchez et al. 2017).

Assessing the presence and extent of carry-over effects can be challenging because of the long timescales and vast distances over which individuals travel after leaving the tropics. Subtle carry-over effects could exist and have an important impact on population dynamics but are challenging to document (Imlay et al. 2019). Current tracking technology for small birds is limited to archival devices (those for which a bird must be recaptured in order to retrieve the information). It is thus only possible to measure carry-over effects for individuals who survive their entire

round-trip migration, a fact which, in itself, may distort estimates. Further, linkages between breeding success and former wintering habitat can only be assessed indirectly through analysis of stable isotopes, which requires a strong moisture gradient across habitats to detect a signal. Here, too, only survivors who arrive at the breeding site can be sampled. Given these challenges, assessment of carry-over effects in a conservation context could instead focus on the underlying physiological mechanisms and signatures that drive population change (O'Connor and Cooke 2015). Direct assessment of physiological measures of stress and body condition could be used in short-term and cost-effective field studies (e.g., Marra and Holberton 1998) and lab studies (Skrip et al. 2016) as a proxy for forecasting the likely strength of carry-over effects.

Studies of carry-over effects would benefit from the use of multiple intrinsic markers. Imlay et al. (2019) integrated stable isotopes, corticosterone levels, telomere dynamics, and body condition of 3 declining migratory species, Bank Swallow (*Riparia riparia*), Barn Swallow, and Cliff Swallow (*Petrochelidon pyrrhonota*), to show that environmental conditions in South American winter habitat resulted in carry-over effects. These included lower annual reproductive success through fewer surviving young, a lower incidence of double brooding, and poorer body condition for those overwintering on some habitats. This study highlights how the multi-marker analysis was vital for inferring the presence of carry-over effects undetectable using a single marker.

Migration Stopover

Although most migration stopover regions and their habitats in the Neotropics are not well identified (Bayly et al. 2018), researchers are beginning to understand how habitat loss or conversion can negatively affect migratory behavior (Bayly et al. 2019). This theme becomes particularly relevant when one considers how some regions and the associated stopover habitats concentrate much of the global population of certain species into areas a fraction of the size of their breeding or wintering grounds. This occurs for Gray-cheeked Thrush in Colombia's Sierra Nevada de Santa Marta (Gómez et al. 2019) and in the Darién region of Panama (Cárdenas-Ortiz et al. 2017, 2020), the latter funneling birds from a region spanning >5,000 km east to west in North America into a narrow corridor barely 100 km wide. The loss of geographically restricted migration/stopover habitats could therefore have a disproportionate effect on some species. In addition, the utility of some stopover sites is still coming into focus. Pyle et al. (2018) studied the postbreeding movements of 140 species captured at more than 900 Monitoring Avian Productivity and Survivorship (MAPS) stations over a 25-yr period to investigate latitudinal, longitudinal, and elevational shifts

from breeding to molting grounds. They found evidence for postbreeding movements to molting areas (many of which were migratory stopover sites) among a variety of North American landbirds, including many migratory species previously assumed to molt on breeding grounds. Movements were heterogeneous, both within and among species, and included areas in the Neotropics; the monsoon region of northwestern Mexico is particularly important in this regard. Logistical challenges notwithstanding, the role of migration stopover sites in the full annual cycle of Nearctic–Neotropical migratory birds merits considerably more attention than it has received to date.

Wintering Habitat Dynamics

Because different breeding or wintering populations do not face the same pressures, and neither increase nor decline at the same rates (Wilson et al. 2018), population-specific information is needed to better inform conservation efforts. Genetically derived population delineations and newly described migration routes are contributing to our understanding of migratory connectivity and habitat use by specific populations of many species (Ruegg et al. 2016). For example, the Willow Flycatcher (*Empidonax traillii*) is a relatively common, widely distributed species across its North American breeding range, although the *extimus* subspecies, which nests in northwestern Mexico and the southwestern United States, is U.S. federally Endangered. Recent work is homing in on the wintering areas used by the breeding population of *extimus* (Ruegg et al. 2020), which would enable conservation strategies to be focused where they are most needed. Similar work using stable isotopes has revealed how wintering Canada Warbler populations are structured across the tropical Andes of Colombia (Gonzalez et al. 2016), making it possible to identify deforestation in the eastern Andes of Colombia as one of the main drivers of declines occurring in eastern but not western breeding populations of this species (Wilson et al. 2018). An isotopic study of the likely breeding origins of Golden-winged Warbler (*Vermivora chrysoptera*) captured throughout their wintering range (Hobson et al. 2016) revealed how steeply declining breeding populations are migrating largely to mountainous regions of northern Colombia and Venezuela. Dramatic declines of Appalachian populations of this species have been linked to disproportionate forest loss on this population's wintering grounds in northern South America (Kramer et al. 2018).

An expanding topic of research is how intraspecific habitat use should be taken into account when evaluating habitat quality and designing appropriate conservation strategies. In many species, males tend to outcompete females or first-year birds in certain habitats (Latta and

Faaborg 2002, Studds and Marra 2005, Wunderle et al. 2014), something that can lead to differential carry-over effects (Marra and Holmes 2001, Drake et al. 2013), while also raising the question about what types of habitat are being conserved. In many species, a proportion of individuals will remain on the same territory throughout the winter, even returning to the same territory in subsequent years (e.g., see Koronkiewicz et al. 2006), though one striking pattern that has emerged from recent wintering studies is the frequency of prolonged stopover behaviors or the use of multiple wintering grounds for some species. In a review of geolocator studies, McKinnon et al. (2013) found that 6 of 11 species studied used secondary wintering areas. Other studies have shown that migrants frequently move among habitats in winter, indicating that behavioral transience is more frequent than previously appreciated (Gómez and Bayly 2010, Peele et al. 2015, Ruiz Gutierrez et al. 2016). Some habitats show an increase in migrant density through the winter, others a decrease (Hilty 1980, Johnson 1980, Lefebvre et al. 1992, Gómez and Bayly 2010), suggesting that individuals possess the capacity to track seasonal changes in food or other resources. Perhaps the species that best exemplifies the range of movement patterns during the nonbreeding season is Bobolink (*Dolichonyx oryzivorus*). Renfrew et al. (2011) noted that "...movements during the...annual cycle of the Bobolink... can be characterized as a continuum, with occasional stops lasting no more than two months". Bobolinks exhibit prolonged periods of stasis in Venezuela and Colombia, which appear to provide fuel for migratory flights both north and south, while stationary periods in Bolivia and Argentina are closest to the traditional concept of wintering areas, but are also used for undergoing an extensive prealternate molt before returning northward (Pyle 1997, Renfrew et al. 2011).

Despite the variation in observed behaviors, similar pressures can be expected to be acting on habitat selection, such as resource abundance (Somveille et al. 2018), relative "safety" (Cuadrado 1997), and prior knowledge (Forrester 2015). Because resources fluctuate with precipitation, moisture gradients may strongly influence the choice of wintering sites (Smith et al. 2010). The timing and seasonality of precipitation should be accounted for, too (Studds and Marra 2007, Rockwell et al. 2017). However, birds wintering in, for example, tropical dry forests may be selecting attributes not directly related to moisture, such as seasonality in resource abundance. Indeed, net annual precipitation by itself may not explain variation in site use, with seasonality in rainfall patterns potentially playing an important role.

Site selection may also be influenced by evolutionary pressures, such as niche conservatism, the tendency of species to retain certain ecological characteristics or behaviors. In a novel study comparing environmental

niche breadth (the 2-dimensional space defined by a set of environmental variables such as mean temperature and average precipitation) between the breeding and nonbreeding periods in Wood Warblers, Gomez et al. (2016) found that migratory warblers were more likely than tropical resident species to track a niche across the year. This suggests that site selection is likely constrained in many species and may also explain why so many Nearctic–Neotropical migratory birds converge on the highlands of Central and South America during the winter (Robinson et al. 1995), where they experience temperatures and precipitation similar to their breeding grounds. Even during migration, some species appear to track the conditions of their environmental niche. For example, Cerulean Warblers use stopover habitats in montane forests in Central America, which tend to mirror many environmental conditions encountered on their breeding grounds (Welton et al. 2012).

On the breeding grounds, most migratory species exhibit high territoriality and between-year site fidelity. A mixture of year-to-year fidelity and territoriality can be advantageous in reducing food search costs (Stamps et al. 2005) and increasing the effectiveness of acquiring food (Greenwood et al. 1982). Given the investments in time and energy required to obtain and hold a winter territory, species with multiple and widely separated wintering sites may not be truly territorial. Here, we reviewed multiple winter territoriality studies for 18 species/subspecies (Table 1). Of those, 7 used multiple wintering areas and, of these 7, 4 had no winter territorial behavior, and 3 had unknown winter territorial behavior. However, high site fidelity has been described even in non-territorial species. In the typically "non-territorial" Prothonotary Warbler (*Protonotaria citrea*), a marked population wintering in coastal scrub in northeastern Costa Rica yielded an estimated year-to-year recapture probability of 0.42 (Wolfe et al. 2013) whereas highly territorial species such as the Willow Flycatcher show an average winter site fidelity of ~68% (Koronkiewicz et al. 2006). It appears that some species may switch modes of territoriality within or between seasons and changes in territoriality between seasons coincide with a switch from an insect-based diet on the breeding grounds to a fruit-based diet on the wintering grounds (Wolfe et al. 2014).

Interspecific competition also plays a role in tropical habitat selection by Nearctic–Neotropical migrants, and may contribute to population limitation, although this topic is relatively poorly understood. Most studies of migratory populations have focused on single species, which tends to neglect migrant–migrant or migrant–resident interactions. The breeding currency hypothesis (Greenberg 1995), supported empirically by Johnson et al. (2005, 2006b), emphasizes that competition from resident birds may have selected for migrants to winter in relatively

TABLE 1. Nearctic–Neotropical migrant species where the use of multiple winter sites or winter territoriality has been documented.

Species	Multiple winter sites	Winter territoriality
Purple Martin (<i>Progne subis</i>)	Yes ¹	No ²
Tree Swallow (<i>Tachycineta bicolor</i>)	No ³	No ⁴
Barn Swallow (<i>Hirundo rustica</i>)	No ⁵	No ⁶
Red-eyed Vireo (<i>Vireo olivaceus</i>)	No ⁸	No ⁷
Bobolink (<i>Dolichonyx oryzivorus</i>)	Yes ¹⁰	No ⁹
Veery (<i>Catharus fuscescens</i>)	Yes ¹²	Unknown ¹¹
Swainson's Thrush (coastal; <i>C. ustulatus</i>)	Yes ¹⁴	Unknown ¹³
Swainson's Thrush (inland; <i>C. ustulatus</i>)	No ¹⁴	Unknown ¹³
Eastern Kingbird (<i>Tyrannus tyrannus</i>)	Yes ¹⁶	No ¹⁵
Western Kingbird (<i>T. verticalis</i>)	Yes ¹⁸	Unknown ¹⁷
Scissor-tailed Flycatcher (<i>T. forficatus</i>)	No ²⁰	No ¹⁹
Gray Catbird (<i>Dumetella carolinensis</i>)	No ²²	Yes ²¹
Golden-winged Warbler (<i>Vermivora chrysoptera</i>)	No ²⁴	Yes ²³
Kirtland's Warbler (<i>Setophaga kirtlandii</i>)	No ²⁶	Yes ²⁵
Prothonotary Warbler (<i>Protonotaria citrea</i>)	Yes ²⁸	No ²⁷
Ovenbird (<i>Seiurus aurocapilla</i>)	No ³⁰	Yes ²⁹
Wood Thrush (<i>Hylocichla mustelina</i>)	No ³²	Yes ³¹
Western Yellow-billed Cuckoo (<i>Coccyzus americanus occidentalis</i>)	No ³⁴	Unknown ³³

References: 1. Fraser et al. (2012), 2. Hill (1993), 3. Bradley et al. (2014), 4. Elliott (1939), 5. Hobson et al. (2015a), 6. Bing (1993), 7. Callo et al. (2013), 8. Cimprich et al. (2020), 9. McKinnon et al. (2013), 10. Renfrew et al. (2020), 11. Hobson and Kardynal (2015), 12. Heckscher et al. (2020), 13. Delmore et al. (2012), 14. Mack and Yong (2020), 15. Cormier et al. (2013), 16. Fitzpatrick (1980), 17. Jahn et al. (2013), 18. Gamble and Bergin (2020), 19. Jahn et al. (2013), 20. Regosin (2020), 21. Ryder et al. (2011), 22. Rappole and Warner (1980), 23. Kramer et al. (2017), 24. J. D. Wolfe personal observations, 25. Cooper et al. (2017), 26. Sykes and Clench (1998), 27. Wolfe and Johnson (2015), 28. Post (1978), 29. Hallworth et al. (2015), 30. Faaborg and Arendt (1984), 31. Stanley et al. (2012), 32. Evans et al. (2020), 33. Sechrist et al. (2012), 34. Hughes (2020).

poor habitats (e.g., migrants leave niches empty for part of the year; Johnson and Sherry 2001). A study on “morphological space” (ecological space as defined by morphological niches and traits) in the Mexican Neotropics by Malpica et al. (2017) supports this hypothesis using seasonal changes in morphological niche-packing and niche-volume.

Breeding Season vs. Wintering Season Limitation

A largely unanswered question for most Nearctic–Neotropical migratory species and populations is what annual cycle phase (breeding, wintering, or migration) is the strongest driver of population change. This gap in knowledge may impact conservation planning, as any efforts to improve habitat in the less limiting season will have little effect in the more limiting season (Sherry and Holmes 1995). Several studies have provided evidence for either breeding or wintering season limitation for certain species. Rappole et al. (2003) argued that, for the U.S. federally Endangered Golden-cheeked Warbler (*Setophaga chrysoparia*), based on both the availability and use of breeding and nonbreeding habitat, nonbreeding habitat in Mexico and Central America is likely more limited than breeding habitat in Texas. Using migratory networks and population modeling, Taylor and Stutchbury (2016) argued that Wood Thrush is more limited by quantity of habitat in winter than in summer. A full life-cycle model for Canada Warbler implicated habitat loss in the tropical eastern Andes as the primary driver of population

declines (Wilson et al. 2018). DeSante et al. (2018) found that population change for Golden-crowned Kinglet (*Regulus satrapa*) was driven primarily by decreases in adult survival in nonbreeding areas, and population changes for “Western” Flycatcher (*Empidonax difficilis/occidentalis*) and Warbling Vireo (*Vireo gilvus*) were driven primarily by changes to first-year survival of young during the nonbreeding season. By contrast, Sherry et al. (2015) argue that, for American Redstart, both winter and summer could be simultaneously driving population change. DeSante et al. (2018) had similar findings for Orange-crowned Warbler (*Oreothlypis celata*). DeSante et al. (2015) conducted one of the largest scale studies to address this question, using 15 yr of demographic data from the MAPS program to examine the demographic drivers of population change for 158 species of passerines and near passerines. Wintering habitat was the primary driving factor in population change for many, if not most, of these species.

Climate Change and Habitat Loss

Impacts to Neotropical habitats from climate change, a topic of enormous importance and breadth, is covered only briefly here. Although some impacts to Nearctic–Neotropical migrant and resident bird species are relatively easy to predict (e.g., for species with limited ranges and narrow habitat requirements), many are unforeseeable because of, for example, the many migratory species whose habitat needs on their wintering grounds are still largely

unknown. As temperature and rainfall patterns shift, habitats will inevitably change, and trophic interactions of primary producers, and secondary and tertiary consumers, will be disrupted; it is therefore likely that many species now considered to have stable populations will become vulnerable.

Under most climate change scenarios, significant areas of habitat for migratory landbirds, including forests, grasslands, and shrublands, will be greatly altered. In some models, habitat loss may increase bird extinctions caused by climate change by 50% and the expected 3.5°C surface warming by the year 2100 may result in 600–900 extinctions of landbird species, 89% of which occur in the tropics (Şekercioglu et al. 2002). In one of the most extensive studies to date, Langham et al. (2015) described the habitat-based climate limitations (“climate envelopes”) for 588 North American species, hundreds of them Nearctic–Neotropical migrants, pinpointing the range of temperatures, rainfall, and other climate characteristics of their habitats. Results indicate that 314 species (53%) will lose more than 50% of their current climatic range by 2080. Although the models examined habitats almost exclusively in North America (plans call for expanding the study to the Neotropics), nonbreeding range in tropical Mexico for Baird’s Sparrow (*Ammodramus bairdii*), for example, is predicted to shrink by 97% by 2050. In addition, the greatly constricted geography of the Central American Isthmus may amplify any habitat losses due to climate change.

Most climate change scenarios forecast increased frequency and strength of storms such as hurricanes and typhoons, a prediction that was bolstered in 2017 when, within a few weeks, hurricanes Irma, Jose, and Maria battered the Caribbean and southeastern U.S. devastating human lives and homes, but also habitat for resident and migratory birds (Wunderle Jr. 2017), as much of the impact occurred during fall migration and early stopover and wintering ground arrival. While many migratory species may sense falling air pressure that precedes a large storm and alter their migration patterns (Dänhardt and Lindström 2001), others cannot and may suffer significant mortality (Dionne et al. 2008). The impacts are still being assessed, but it is already known that the effects on nonbreeding habitats for many Nearctic–Neotropical migrants were widespread and severe.

One of the first studies that identified the potential for the El Niño–Southern Oscillation (ENSO) to impact migrant birds by affecting their Neotropical habitats was conducted by Sillett et al. (2000), who found a correlation between El Niño events in Jamaica, which are associated with droughts in the Caribbean Basin, and reductions in Black-throated Blue Warbler (*Setophaga caerulea*) annual adult survival, presumably because these droughts impact the availability of food needed to survive the winter and to fuel spring migration to

breeding grounds. Subsequently, Wilson et al. (2011) found support for negative impacts of Caribbean drought, especially centered on Cuba, on Breeding Bird Survey population trends in eastern North America in American Redstarts. Similar effects of winter season drought in the Caribbean have been suggested for Ovenbirds (*Seiurus aurocapila*; Brown and Sherry 2006), Northern Waterthrush (*Seiurus noveboracensis*; Smith et al. 2010), and the endangered Kirtland’s Warbler (Wunderle and Arendt 2017). For migratory species from western North America, Lamanna et al. (2012) found that annual survival of Swainson’s Thrush was strongly influenced by ENSO-related weather changes during one or more periods of its annual cycle. The clear message from this body of research is that ENSO-related droughts impact habitats, and therefore populations, in a variety of birds, probably via food supply. Droughts will likely increase in magnitude with intensification of climate change.

Phenological Mismatch

Phenology, the timing of seasonal life-cycle events, is fundamental to species’ ecology and life history, providing a valuable indicator of environmental impacts on ecosystems (IPCC 2014). Climate-driven changes in habitat phenology can impact migratory bird populations through disruption or “mismatch” of previously synchronized trophic interactions (Sherry and Holmes 1996). Mismatch in habitat conditions between Neotropical wintering and stopover habitats may impact wintering body condition, overwinter survival, and spring departure (Marra and Holberton 1998, Strong and Sherry 2000) with carry-over effects on reproductive success (Norris et al. 2004) and migration timing (Paxton and Moore 2015) for migrating birds. For example, Neotropical habitat quality has been associated with an overlap of wintering migrants with invertebrate prey abundance (Strong and Sherry 2000), which is influenced by winter precipitation (McKellar et al. 2013, Rockwell et al. 2017), highlighting that a predictable or seasonally stable optimal window plays an important role in bird movements and habitat selection on the wintering grounds (Parrish and Sherry 1994, Kresnik and Stutchbury 2014).

Phenological information from the Neotropics is not representative of the large land area and biodiversity of the region (Mendoza et al. 2017, Merrick et al. 2019), despite the fact that most climate projections suggest that tropical biomes will experience among the greatest increases in extreme heat and precipitation events (Garcia et al. 2014), with concurrent changes to the abundance and phenological timing of pollen, fruits, seeds, and insects expected in these habitats as well.

Mismatch may be most significant for long-distance migrants (La Sorte and Fink 2017), especially species which

depend on highly seasonal food resources that have a narrow temporal distribution of availability and potential for “match” (Both et al. 2010), as well as species with more restricted diets (Mallord et al. 2017). Research on wintering habitat ecology has revealed interactions of habitat type and habitat structure, food availability, and bird movements, although research is lacking on long-term distributional patterns that could reveal phenological mismatches. Age- and sex-related habitat segregation within wintering bird populations adds complexity to the task of fully understanding the implications of phenological mismatch (Parrish and Sherry 1994, Kresnik and Stutchbury 2014). Although there is significant variation in phenotypic plasticity for responding to interannual variation and trends, there are likely distributional limits to most populations and species (Fraser et al. 2019), which, if trends continue, may soon be exceeded. Links between variation in local and hemispheric weather and climate factors, trophic-associated habitat quality, and avian responses should be considered within the context of phenological mismatch across species assemblages and inherent latitudinal differences in the ecological processes and mechanisms involved.

Priority Habitats

The preceding sections explained why tropical habitats are so pivotal in the full annual cycle of migratory birds, while hinting at the finding that some vegetation types occupied by migrants may be of greater importance than others. We still have much to learn about the relative use and quality of tropical habitats, but our current knowledge is sufficient to draw attention to specific habitats that are both important to declining species and are themselves under threat. Multiple studies have converged on the conclusion that humid pre-montane forests in the tropical Andes provide habitat for a suite of steeply declining species including Cerulean Warbler, Canada Warbler, Golden-winged Warbler, and Olive-sided Flycatcher (Colorado et al. 2012, Kramer et al. 2017, Céspedes and Bayly 2018, Wilson et al. 2018). The tropical Andes span a broad elevation range, with multiple peaks above 5,000 m; by contrast, migratory birds occupy a relatively restricted elevation band primarily between 1,000 and 2,000 m. These intermediate elevations have undergone extreme transformations (Rodríguez-Eraso et al. 2013), making way for coffee plantations, cattle pastures, and human settlements (Correa Ayram et al. 2015). It is no surprise then that Andean forests in Venezuela, Colombia, and, to a lesser extent, Ecuador and Peru, are considered a priority for the conservation of migratory landbirds. In addition, the relative impacts of habitat loss from climate change (discussed below) may be greatly amplified in such areas.

In Central America, the forested highlands have also been recognized as critical habitat for several species. Humid broadleaf forests in mountainous regions between

500 and 2,500 m of Costa Rica, Panama, Nicaragua, and Honduras are habitat for Golden-winged Warblers (Bennett et al. 2016), while highland pine-oak forests (1,100–2,400 m) in northern Central America are occupied by Golden-cheeked Warblers (Rappole et al. 2000). Bicknell's Thrush (*Catharus bicknelli*) also winter in humid montane broadleaf forests in the Greater Antilles (McFarland et al. 2013). While a disproportionate number of declining and threatened species are associated with montane forest, lowland habitats also have a role. Humid lowland broadleaf forests in Central America provide high quality habitat for Wood Thrush and Kentucky Warbler (*Geothlypis formosa*), among others, while mangroves, especially black mangrove (*Avicennia germinans*) stands, hold a high diversity and richness of migrants in Central America (Gómez and Bayly 2010) and the Caribbean. In northern South America, occupancy, emigration, and extinction rates all point to mangroves and wooded wetlands as vital habitats for the declining Prothonotary Warbler (Bulluck et al. 2019).

Studies examining the factors driving presence/abundance, as well as comparing habitat quality, reveal a common pattern among several migratory species, an increasing probability of occurrence or abundance in more humid and more mature forests (Johnson et al. 2006a, McFarland et al. 2013, Ruiz-Sanchez et al. 2017, Céspedes and Bayly 2018). For example, increasing canopy height not only results in higher densities of Canada Warblers in the Andes (Céspedes and Bayly 2018) but also a lower rate of extinction from the wetland habitats occupied by Prothonotary Warblers (Bulluck et al. 2019). It follows that the loss of mature forests is likely to negatively impact a range of species. Nonetheless, many migratory landbirds will occupy transformed tropical habitats and are not forest-dependent. Shade-coffee plantations, for example, are widely recognized as a better alternative to many other possible land uses (Komar 2006, González-Prieto 2018a) and may even support similar densities of migrants as the forests they replaced (Bakermans et al. 2009).

How transformed tropical habitats compare with natural habitats is a topic that has received relatively limited attention (González-Prieto 2018b), despite its considerable importance when it comes to deciding between land-sharing or land-sparing approaches (Chandler et al. 2013) for migratory bird conservation. For American Redstarts wintering in Jamaica, coastal mangroves are of higher quality than either shade coffee or citrus plantations (Johnson et al. 2006a). By contrast, Bakermans et al. (2009) concluded that shade coffee was of equal or higher quality than nearby primary forest in Venezuela based on the density of migrants. While density alone is not necessarily a reliable measure of habitat quality, the degree of variation in density between shade coffee and forest among different

migrants is noteworthy. Canada Warblers, for example, are more abundant in forest (Céspedes and Bayly 2018), while the mean abundance of Blackburnian Warblers (*Setophaga fusca*) and Tennessee Warblers per mixed species flock was higher in coffee (McDermott and Rodewald 2014). These mixed results highlight the role shade coffee plays as a wintering habitat but it is notable that during migration, when energy needs are elevated, birds departing from or stopping over in forest have been found to fuel faster, attain larger energy reserves, and migrate in longer stages (Bayly et al. 2016, 2019, González-Prieto 2018b).

The quality of other agricultural systems that include shade trees should be further explored, including cocoa (Schroth and Harvey 2007), cardamom, silvo-pastures (McDermott and Rodewald 2014), and mixed native hardwood plantations (Bennett 2018).

The Future of Neotropical Habitats for North American Migrants

The last few decades have underscored the, sometimes contradictory, trends in the study and conservation of North American breeding birds on their stopover and wintering grounds in the Neotropics. Enormous strides have been made in our understanding of movements, population dynamics, habitat use, and successful conservation practices. Tracking devices are becoming smaller and cheaper every year and, in the near future, scientists may have mapped the movements and seasonal habitats of every species that migrates between North America and the tropics. The development of new population models, and the spread of eBird, the Monitoring Overwinter Survival program, and other community science programs, are providing valuable insights into how birds move on their wintering grounds. Concurrently, Latin American and Caribbean governments and homegrown conservation movements have become empowered and capable of successfully carrying out complex and comprehensive research and conservation initiatives, and among the public, the importance of birds and their conservation has grown considerably (see Dayer et al. 2020).

But will all of these improvements lead to better outcomes for the “birds of two worlds”? Many of the threats that have been well documented for decades, such as habitat loss and fragmentation, invasive species, pollution, and urbanization, continue apace and have, in many areas, accelerated. Add to these emerging threats factors such as climate change and the associated impacts of sea level rise and phenological mismatch, and it can be difficult to remain optimistic about the future of Nearctic–Neotropical migratory birds.

If the negative trends are to be reversed and the positive ones preserved, large reserves and protected areas will be more important than ever. But these areas need to be designed with a multitude of factors in mind, such as the

full cooperation of local stakeholders, the complexities of political and social realities, and—crucially—the potential changes to the landscape that climate change will bring. Networks of protected areas need to incorporate extensive topographical diversity, cover wide elevation ranges, have high connectivity, and integrate human-dominated landscapes into conservation schemes, something which is not currently universally done (Langham et al. [2015], for example, found no strong associations between projected climate sensitivities and existing conservation priorities). Many Nearctic–Neotropical migratory bird species vulnerable to climate change are not currently considered threatened with extinction, often due to lack of knowledge, so systematically and regularly gathering information on the ecology and current and future distributions of these species is an urgent matter. One of the goals of the international bird conservation consortium Partners in Flight is “keeping common birds common” (Rosenberg et al. 2016).

In addition, while many countries in the Neotropics have established protected areas, they lack the financial resources to actively manage them, and many remain “paper parks”—lines on a map with no de facto protection—while habitat loss goes on around and sometimes inside their borders. Because a relatively high percentage of land in the Neotropics is privately owned (and therefore not legally protected in the long term), land-sharing systems that incorporate primary forest and a diversity of other habitats will be a key component of future conservation initiatives. Because locally based, long-term tropical bird monitoring and conservation programs based on adaptive management are essential to help protect birds against the myriad threats they face, local conservation efforts need to be bolstered with all the resources cooperating countries and non-governmental organizations can muster. And, because the birds we all care about traverse much of the hemisphere in their annual migrations, cooperation and resources must come from all countries in the hemisphere.

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